

RESEARCH LETTER – Environmental Microbiology

Synergistic effect of crude oil plus dispersant on bacterial community in a Louisiana salt marsh sediment

Mohammed Al-Jawasim, Kewei Yu and Joong-Wook Park*

Department of Biological and Environmental Sciences, Troy University, Troy, AL 36082, USA

*Corresponding author: Department of Biological and Environmental Sciences, Troy University, Troy, AL 36082, USA. Tel: +334-808-6416;

Fax: +334-670-3662; E-mail: jwpark@troy.edu**One sentence summary:** Synergistic effect of crude oil plus dispersant (Corexit 9500A) significantly altered indigenous bacterial communities in a Louisiana salt marsh sediment after 30 days of incubation.

Editor: Yu-Zhong Zhang

ABSTRACT

A combined effect of crude oil plus dispersant (Corexit 9500A) significantly altered indigenous bacterial communities in a Louisiana salt marsh sediment after 30 days of incubation; the crude oil and/or Corexit 9500A treatments triggered shifts in bacterial communities and the shifted bacterial structure by crude oil plus Corexit 9500A was considerably different from those by either crude oil or Corexit 9500A. However, the synergistic effect of crude oil plus Corexit 9500A was not observed after 7 days of incubation; the bacterial community was slightly shifted by Corexit 9500A and the crude oil did not trigger any bacterial community shift after 7 days of incubation. DNA sequencing data indicated that *Chromobacterium* species was enriched in the Corexit 9500A microcosms after 7 days of incubation, while *Pseudomonas*, *Advenella*, *Acidocella* and *Dyella* spp. were enriched after 30 days of incubation. *Parvibaculum* was a dominant species in the crude oil microcosms after 30 days of incubation. *Rhodanobacter*, *Dyella* and *Frateuria* spp. were dominant in crude oil plus Corexit 9500A microcosms after 30 days of incubation. Our data show that the effect of crude oil plus Corexit 9500A on bacterial community is synergistic, and thus the dispersant effect should be considered with the spilled oil to correctly evaluate the environmental impact.

Keywords: crude oil; Corexit; synergistic effect; dispersed oil; bacterial community; salt marsh

INTRODUCTION

Dispersant application is one of the strategies to mitigate oil spill impacts (Fiocco and Lewis 1999) that has been applied since the 1950s (Ramachandran et al. 2004). Because of the amphipathic nature of surfactants (Canevari 1969; Fiocco and Lewis 1999), the dispersants arrange themselves at oil–water interfaces to minimize the interfacial tension and break oil slicks into tiny droplets that settle in the water column (Goodbody-Gringley et al. 2013; Prince and Butler 2013). Consequently, the dispersant application increases biodegradation of crude oil by enhancing oil bioavailability to a large number of bacterial species, partic-

ularly hydrocarbon-degrading bacteria (Fiocco and Lewis 1999; Prince, Lessard and Clark 2003). Varadaraj et al. (1995) reported that Corexit 9500A—a dispersant used extensively in the Deepwater Horizon oil spill—positively affected the microbial growth and enhanced the rate of microbial oil degradation, since sorbitan in Corexit 9500A can be served as a nutrient for microbial growth.

On the other hand, a substantial amount of research has demonstrated the dispersant toxicity. Singer et al. (1996) showed that Corexit 9500A can elicit acute toxic effects to aquatic organisms such as kelp forest mysid and red abalone. Hamdan and Fulmer (2011) demonstrated that reproduction and viability

Received: 18 June 2015; Accepted: 24 August 2015

© FEMS 2015. All rights reserved. For permissions, please e-mail: journals.permissions@oup.com

of two oil-degrading bacterial isolates were significantly inhibited by Corexit 9500A. Besides the toxicity of dispersants themselves, toxic effects of dispersed oil or chemically enhanced water-accommodated fraction (CEWAF) has been extensively studied (Ramachandran *et al.* 2004; Place *et al.* 2010). A considerable amount of research supports that CEWAF is toxic to fish (Ramachandran *et al.* 2004, 2006; Gardiner *et al.* 2013; Rico-Martinez, Snell and Shearer 2013; Adams, Swezey and Hodson 2014; Dussauze *et al.* 2015), copepod (Gardiner *et al.* 2013; Lee *et al.* 2013; Cohen, McCormick and Burkhardt 2014), crab (Chase *et al.* 2013), shrimp/abalone (Rico-Martinez, Snell and Shearer 2013), sea urchin embryo (Rial, Vazquez and Murado 2014), coral larva (Goodbody-Gringley *et al.* 2013), rotifera (Singer *et al.* 1998) and diatom (Hook & Osborn, 2012). However, there still are controversial issues on synergistic toxicity of CEWAF as compared to non-dispersed oil. Some researchers insisted no synergistic toxicity of CEWAF (Adams, Swezey and Hodson 2014; Dussauze *et al.* 2015), while others claimed synergistic toxic effect of oil and dispersant (Singer *et al.* 1998; Rico-Martinez, Snell and Shearer 2013).

Various eukaryotes were used as model organisms to evaluate the synergistic effect of CEWAF, but study is still necessary to investigate the synergistic effect of CEWAF on microbial community. There are several reports to show the effect of CEWAF on microorganisms (Ortmann *et al.* 2012; Baek, Son and Shim 2013), but they did not examine whether the synergistic effect exists. Considering the crucial role of bacteria in marine food webs (Ortmann *et al.* 2012), it is important to investigate the existence of synergistic effect of dispersed oil on marine bacteria. Our research objective is to investigate the synergistic effect of crude oil plus Corexit 9500A on bacteria in a Louisiana salt marsh sediment. Bacterial community shifts after expose to crude oil and/or Corexit 9500A were monitored after 7 and 30 days of incubation.

MATERIALS AND METHODS

Sampling area

A sediment sample was collected in May 2013 from a salt marsh at the eastern side of Lake Pontchartrain, Louisiana (N30° 08.782' W89° 44.665'). The site was dominated by two marsh plants, *Spartina alterniflora* and *S. patens*. After eliminating the surface vegetation, the top 30 cm sediments were collected, stored in sterilized containers, transported on ice to the laboratory and stored at 4°C before use.

Chemical analysis

The sediment sample was sent to the Central Analytical Instruments Research Laboratory, Louisiana State University (Baton Rouge, LA) for chemical analysis. EPA methods 200.7, 300.0 and 365.3 were used to analyze metals, anions and total phosphorous, respectively. Chemical analyses of the sediment were performed to assess factors that might affect the indigenous bacterial community (Table S1, Supporting Information).

Microcosm setup and DNA extraction

West Texas Intermediate (WTI), also known as Texas Light Sweet crude oil, was purchased from Texas Raw Crude (Midland, TX) and the dispersant Corexit 9500A was generously provided by the Nalco Energy Services (Sugar Land, TX). Four sets of microcosms were (1) untreated control, (2) microcosms with 0.2% (v/w) of Corexit 9500A, (3) microcosms with 2% (v/w) of WTI and (4) microcosms with 0.2% of Corexit 9500A and 2% of WTI. The micro-

cosms were set up in triplicate. One gram of sediment was aseptically transferred to 2 mL sterile tubes. Corexit 9500A and/or WTI were added via micropipettes and mixed with the sediment using a tip on a micropipette. The microcosm tubes were sealed with a Teflon-coated cap and incubated without shaking at 30°C. Since approximately two-thirds of air-filled headspace was formed in the microcosm tubes, sediments were incubated under aerobic conditions. The microcosms were sampled after 7 and 30 days of incubation. Total DNA was extracted from 0.3 g of sediment using the PowerSoil DNA Isolation Kit (MoBio Laboratories, Carlsbad, CA) and the DNAs were stored at -20°C prior to analysis.

Polymerase chain reaction (PCR)

Nested PCR method was performed to amplify the bacterial 16S rRNA genes. The primers 27F and 1522R (Giovannoni 1991) were used for a first-round PCR to amplify the entire bacterial 16S rRNA genes. The final volume of the PCR was 50 µL containing 10 pmol of primers, 1 µL of template DNA, 0.25 mM of dNTP, 5 µL of 10x PCR buffer and 1 U of Green Taq DNA polymerase (GenScript, Piscataway, NJ). PCR was carried out with a DNA thermal cycler (GeneAmp PCR System 2700, Applied Biosystems, Foster City, CA) at an initial temperature of 94°C for 5 min, followed by 30 cycles of 94°C for 20 s, 55°C for 45 s and 72°C for 45 s. A final elongation step was 72°C for 7 min. In the second-round PCR, the primer 341F with GC clamp and 534R were used to amplify the V3 region of the bacterial 16S rRNA gene (Muyzer, De Waal and Uitterlinden 1993). Approximately 1 µL of the first-round PCR product was used as a template for a second-round PCR. PCR conditions and constituents are the same as described above. The resulting PCR products were confirmed by an agarose gel electrophoresis.

Denaturing gradient gel electrophoresis (DGGE) and image analysis

The second-round PCR products were separated on an 8% polyacrylamide gel with a 40–60% denaturing gradient of urea in 1.0x TAE buffer by a Bio-Rad DCode Universal Mutation Detection System (Bio-Rad Laboratories, Hercules, CA). Approximately 45 µL of PCR products were loaded and the electrophoresis was conducted at 40 V for 15 h. After electrophoresis, gels were stained with ethidium bromide and photographed on a UV transilluminator (Fisher Scientific, Pittsburgh, PA). DGGE images were analyzed by the PyElph software (version 1.4) to construct phylogenetic trees based on DGGE profiles using the unweighted pair group method with arithmetic mean (UPGMA) algorithm.

DNA sequencing and data analysis

The selected DGGE bands were excised using a sterilized blade and incubated with 50 µL of distilled water at 4°C overnight. PCR was carried out to reamplify the eluted DNA with the PCR conditions as described above except that the PCR reaction was performed for 35 cycles. The DGGE analysis was conducted to verify the purity and position of the re-amplified DNAs. If necessary, the DGGE bands were reexcised and the process of PCR-DGGE was repeated until a single DGGE band was confirmed. The resulting PCR products were purified with the MEGAquick-spin Total Kit (iNTRON Biotechnology Inc., Korea) for DNA sequencing. The DNA sequencing was outsourced to the Genewiz Inc. (Genewiz Inc., South Plainfield, NJ). Each sample was sequenced with forward and reverse primers separately. The DNA sequences were assessed using the Chromas Lite Ver. 2.1.1

(Mullen et al. 2011), and the basic local alignment search tool (BLAST) algorithm was used to search homologous sequences in the GenBank DNA libraries (Altschul et al. 1990).

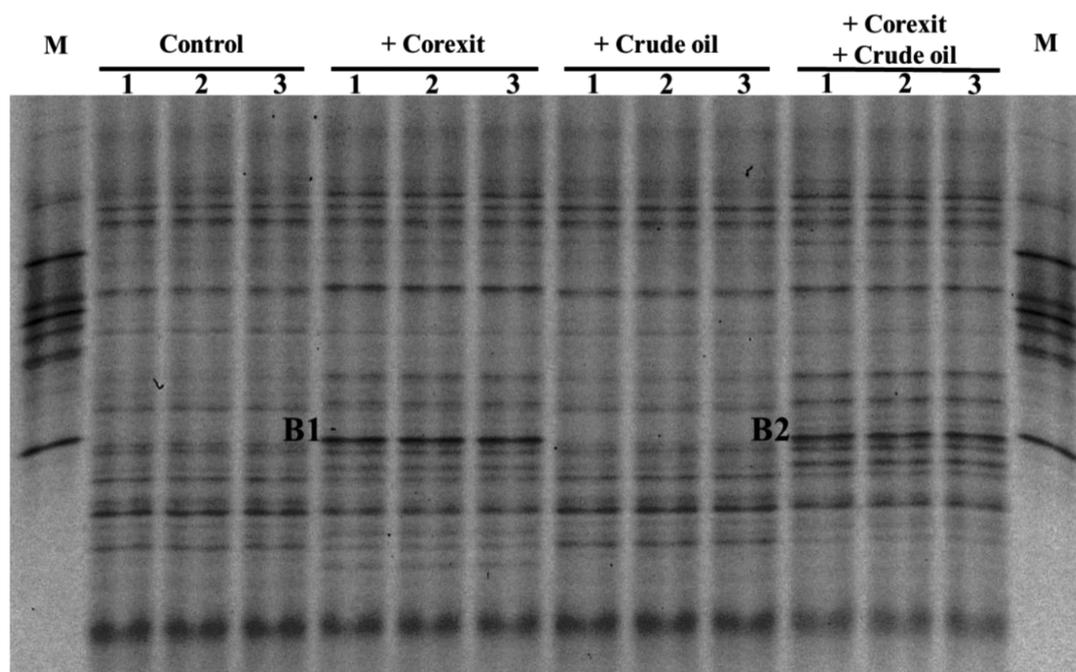
RESULTS AND DISCUSSION

Effect of crude oil and/or Corexit 9500A after 7 days of incubation

In terms of patterns and intensity of DGGE bands, control samples were almost the same as crude oil treatments (Fig. 1A). The

same result was again confirmed by the PyElph analysis (Fig. 1B), which strongly suggests that crude oil treatments have no effect on the bacterial community after 7 days of incubation. Consistent with our observations, it is reported that bacterial abundance and metabolic activity in microcosms were not significantly affected by diesel containing 0.55–55 ppm (dry weight) of polycyclic aromatic hydrocarbons (PAHs) after 7 days of exposure at any concentration tested (Carman, Means and Pomarico 1996). Microbial communities need an adaptation period after they are exposed to a new substrate, which may explain no effect of crude oil on the bacterial community.

(A)



(B)

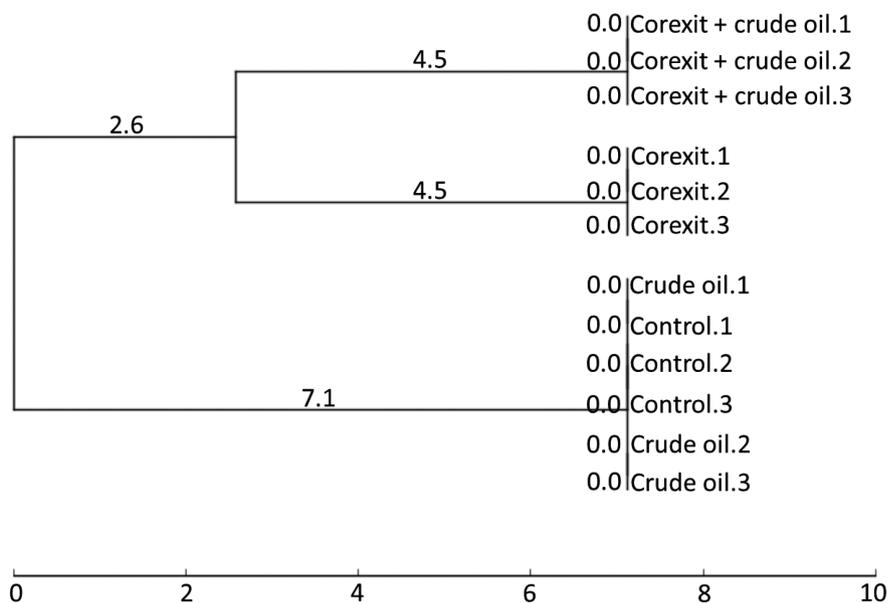


Figure 1. Bacterial community shift after 7 days of incubation. (A) DGGE profiles of bacterial 16S rRNA gene fragments in sediment microcosms with 0.2% of Corexit 9500A, 2% of crude oil and 0.2% of Corexit 9500A plus 2% of crude oil. (B) UPGMA dendrogram of DGGE profiles. The values on the horizontal lines stand for genetic distances among treatments in percentages. M: custom marker; numbers 1, 2 and 3 represent individual triplicates.

Table 1. Close relatives of major DGGE bands.

Band	Close relatives in GenBank databases	GenBank Accession no	% Identity
B1	<i>Chromobacterium violaceum</i> strain ATCC 12472 ^T	NR_074222	92.74
B2	<i>Chromobacterium violaceum</i> strain ATCC 12472 ^T	NR_074222	92.41
B3	<i>Pseudomonas</i> sp. CZ5	GQ903480	91.19
B4	<i>Adivenella kashmirensis</i> strain W13003	KF147541	76.38
B5	<i>Acidocella</i> sp. PFBC	KC590088	98.51
B6	<i>Dyella ginsengisoli</i> strain LA-4	EF191354	96.86
B7	<i>Dyella ginsengisoli</i> strain LA-4	EF191354	99.37
B10	<i>Parvibaculum lavamentivorans</i> strain DS-1 ^T	NR_074262	97.01
B11	<i>Rhodanobacter denitrificans</i> strain 2APBS1 ^T	NR_102497	97.48
B12	<i>Dyella ginsengisoli</i> strain Gsoil 3046 ^T	NR_041370	99.37
B13	<i>Frateuria</i> sp. WJ64	AY495957	99.37

The DGGE profiles of Corexit 9500A treatments were considerably similar to those of crude oil plus Corexit 9500A treatments (Fig. 1A), which was confirmed by the PyElph analysis (Fig. 1B). These data provide additional support that crude oil treatments have no effect on the bacterial community after 7 days of incubation, since the patterns of crude oil plus Corexit 9500A treatments were affected mainly by the Corexit 9500A. It is also noticeable that the effect of Corexit 9500A on bacterial community was detectable after 7 days of incubation. One major DGGE band at the same position and intensity was observed in Corexit 9500A and crude oil plus Corexit 9500A treatments (B1 and B2, Fig. 1A). DNA sequencing data showed these two DGGE bands were homologous to *Chromobacterium violaceum* strain ATCC 12472^T (Table 1). Itah and Essien (2005) demonstrated that *C. violaceum* can extensively degrade hydrocarbons. Since band B1 was appeared in microcosms treated only with Corexit 9500A, it is considered that band B1 was appeared in response to Corexit 9500A. Considering Corexit 9500A is a mixture of hydrocarbons, glycols and sodium dioctyl sulfosuccinate (Bælum et al. 2012), our observation suggests that *C. violaceum* thrive by metabolizing hydrocarbons of Corexit 9500A.

Synergistic effect of crude oil plus Corexit 9500A after 30 days of incubation

The bacterial community structures were greatly shifted after 30 days of exposure to Corexit 9500A, crude oil or both (Fig. 2). The UPGMA dendrogram demonstrated that crude oil plus Corexit 9500A treatments were the most phylogenetically distinctive cluster (Fig. 2B). Crude oil plus Corexit 9500A treatments had unique DGGE bands patterns and intensity, which were completely different from those of other treatments or control (Fig. 2A). Hence, our data strongly suggest that crude oil plus Corexit 9500A treatments that contain CEWAF synergistically triggers bacterial community shift. It is also noticeable that this synergistic effect was not observed after 7 days of incubation (Fig. 1). The DNA sequence-based analysis demonstrated that bacterial species thrived in crude oil plus Corexit 9500A treatments were phylogenetically different from those in other treatments (Table 1), which also support the presence of synergistic effect of crude oil plus Corexit 9500A.

Phylogenetic analysis of major DGGE bands after 30 days of incubation

DNA sequence analysis showed that the close relatives of major DGGE bands were associated with biodegradation of crude oil,

excluding band B11 which was phylogenetically close to a denitrifying bacterium (Table 1).

The major DGGE bands in Corexit 9500A treatments were bands B3, B4, B5, B6 and B7. The close relative of band B3 was *Pseudomonas* sp. CZ5, a PAH-degrading bacterium isolated from PAH-contaminated sludge samples from a chemical plant in China (Zhou et al. 2013). The close relative of band B4 was hydrocarbon-degrading *Adivenella kashmirensis* strain W13003 isolated from PAH-contaminated marine sediments in China (Wang et al. 2014). The close relative of band B5 was *Acidocella* sp. PFBC, an acidophilic heterotrophic bacterium that can metabolize several aromatic hydrocarbons, such as phenol, benzyl alcohol and benzoate (Jones, Hedrich and Johnson 2013). The close relative of band B6 and B7 was *Dyella ginsengisoli* strain LA-4, which was isolated from an activated sludge of a petrochemical company in China, which can use hydrocarbons as its carbon and energy sources (Li et al. 2009a,b). It is considered that the relatives of above oil-degrading bacteria were found in our microcosms, since the microcosms were setup with the estuarine sediment and incubated under aerobic condition with Corexit 9500A, which contains hydrocarbons (Bælum et al. 2012).

The bands B8, B9 and B10 were present in microcosms treated with crude oil (Fig. 2A). We failed to sequence the bands B8 and B9, but the band B10 was sequenced. Sequence analysis demonstrated that the band B10 was homologous to *Parvibaculum lavamentivorans* strain DS-1^T (Table 1). This strain was isolated from an activated sludge for its ability to degrade linear alkylbenzene sulfonate surfactants (Schleheck et al. 2004, 2007). The genus *Parvibaculum* includes hydrocarbon-degrading species, which were detected in several hydrocarbon-contaminated environments (Alonso-Gutiérrez et al. 2009; Paixão et al. 2010; Wang et al. 2010). Consequently, the enrichment of genus *Parvibaculum* suggests that the hydrocarbon-degrading and/or biosurfactants-producing bacteria may exist in the microcosms treated with crude oil.

The bands B11, B12 and B13 in crude oil plus Corexit 9500A treatments were chosen for DNA sequencing. These bands were phylogenetically homologous to *Rhodanobacter denitrificans* strain 2APBS1^T, *D. ginsengisoli* strain Gsoil 3046^T and *Frateuria* sp. WJ64, respectively (Table 1). *Rhodanobacter denitrificans* strain 2APBS1^T contains nitrate, nitrite, nitric oxide and nitrous oxide reductase genes and is able to perform complete denitrification under anaerobic condition (Green et al. 2010; Prakash et al. 2012). A recent study showed the inhibition effect of Corexit 9500A on denitrification rate after 2 weeks of incubation (Pietroski, White and DeLaune 2015). The other study reported that Tween 80—a major component of Corexit 9500A (Varadaraj et al. 1995)—delayed the

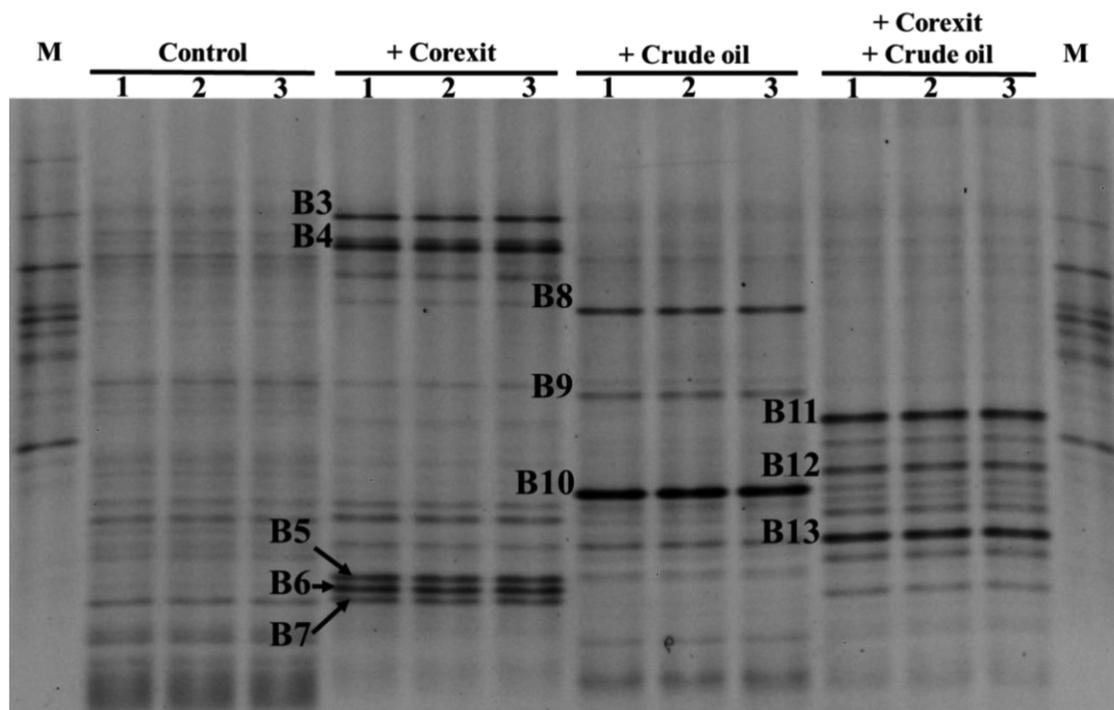
denitrification after 3 weeks of incubation, but the delaying effect of denitrification was not observed and the biodegradation of petroleum hydrocarbons was increased after 6 weeks of incubation (Zhang, Zheng and Lo 2015). Our data suggest the presence of denitrifying bacteria as one of the major species after 30 days of crude oil plus Corexit 9500A treatments (band B11, Fig 2A), which is consistent with previous studies. However, further studies are necessary to confirm this idea. The species *D. ginsengisoli* includes many hydrocarbon-degrading strains such as LA-4 (Li et al. 2009a) and MS2 (Chang, Chang and Yuan 2008). It is interesting to note that the bands B6, B7 and B12 were ap-

peared at different positions, but they were homologous to the same species *D. ginsengisoli* (Table 1). The genus *Frateuria* includes several bacterial species such as *Frateuria* sp. ANA-18 (Mura-kami et al. 2003) and *Frateuria aurantia* (Zemb et al. 2012) that can metabolize hydrocarbons as the sole source of carbon and energy.

CONCLUSION

While a 7-day exposure did not show the synergistic effect of crude oil plus Corexit 9500A, the synergistic effect was observed

(A)



(B)

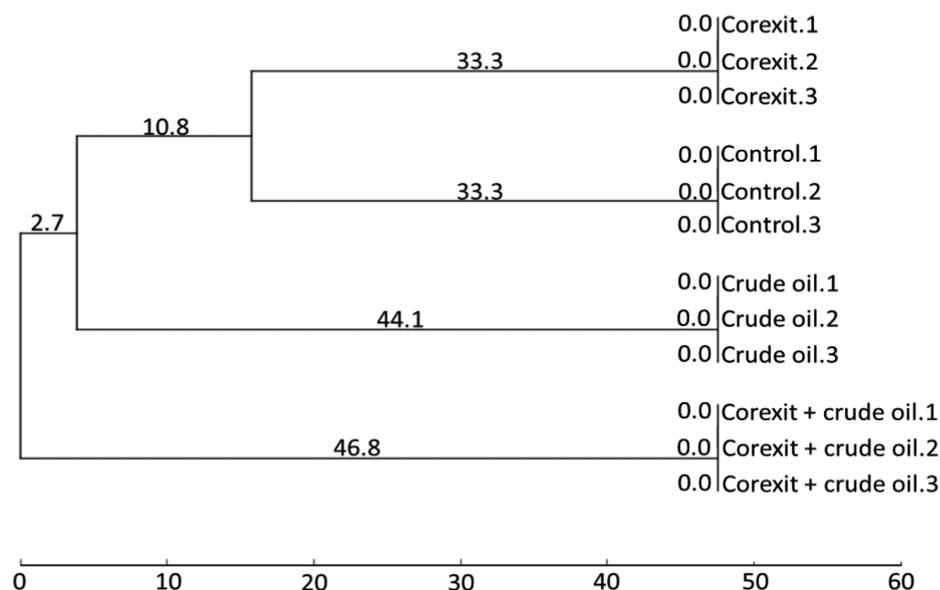


Figure 2. Bacterial community shift after 30 days of incubation. (A) DGGE profiles of bacterial 16S rRNA gene fragments in sediment microcosms with 0.2% of Corexit 9500A, 2% of crude oil and 0.2% of Corexit 9500A plus 2% of crude oil. (B) UPGMA dendrogram of DGGE profiles. The values on the horizontal lines stand for genetic distances among treatments in percentages. M: custom marker; numbers 1, 2 and 3 represent individual triplicates.

after a 30-day exposure. Considering crude oil plus Corexit 9500A contain CEWAF, our data support the presence of synergistic effect of CEWAF on indigenous bacteria in the Louisiana salt marsh sediment. Our findings strongly suggest that the dispersant effect should be considered with the spilled oil to correctly evaluate the environmental impact.

SUPPLEMENTARY DATA

Supplementary data are available at FEMSLE online.

FUNDING

This work was supported by the Higher Committee for Education Development in Iraq (HCED) grant and the Faculty Development Research grants, Troy University, Troy AL.

Conflict of interest. None declared.

REFERENCES

- Adams J, Swezey M, Hodson PV. Oil and oil dispersant do not cause synergistic toxicity to fish embryos. *Environ Toxicol Chem* 2014;**33**:107–14.
- Alonso-Gutiérrez J, Figueras A, Albaigés J, et al. Bacterial communities from shoreline environments (Costa da Morte, North-western Spain) affected by the Prestige oil spill. *Appl Environ Microb* 2009;**75**:3407–18.
- Altschul SF, Gish W, Miller W, et al. Basic local alignment search tool. *J Mol Biol* 1990;**215**:403–10.
- Baek SH, Son M, Shim WJ. Effects of chemically enhanced water-accommodated fraction of Iranian Heavy Crude Oil on periphytic microbial communities in microcosm experiment. *B Environ Contam Tox* 2013;**90**:605–10.
- Bælum J, Borglin S, Chakraborty R, et al. Deep-sea bacteria enriched by oil and dispersant from the Deepwater Horizon spill. *Environ Microbiol* 2012;**14**:2405–16.
- Canevari GP. General dispersant theory. In: *International Oil Spill Conference, American Petroleum Institute*. Washington DC, USA, 1969;171–7.
- Carman KR, Means JC, Pomarico SC. Response of sedimentary bacteria in a Louisiana salt marsh to contamination by diesel fuel. *Aquat Microb Ecol* 1996;**10**:231–41.
- Chang BV, Chang IT, Yuan SY. Biodegradation of phenanthrene and pyrene from mangrove sediment in subtropical Taiwan. *J Environ Sci Heal A* 2008;**43**:233–8.
- Chase DA, Edwards DS, Qin G, et al. Bioaccumulation of petroleum hydrocarbons in fiddler crabs (*Uca minax*) exposed to weathered MC-252 crude oil alone and in mixture with an oil dispersant. *Sci Total Environ* 2013;**444**:121–7.
- Cohen JH, McCormick LR, Burkhardt SM. Effects of dispersant and oil on survival and swimming activity in a marine copepod. *B Environ Contam Tox* 2014;**92**:381–7.
- Dussauze M, Pichavant-Rafini K, Le Floch S, et al. Acute toxicity of chemically and mechanically dispersed crude oil to juvenile sea bass (*Dicentrarchus labrax*): absence of synergistic effects between oil and dispersants. *Environ Toxicol Chem* 2015;**34**:1543–51.
- Fiocco RJ, Lewis A. Oil spill dispersants. *Pure Appl Chem* 1999;**71**:27–42.
- Gardiner WW, Word JQ, Word JD, et al. The acute toxicity of chemically and physically dispersed crude oil to key Arctic species under Arctic conditions during the open water season. *Environ Toxicol Chem* 2013;**32**:2284–300.
- Giovannoni S. The polymerase chain reaction. In: Stackebrandt E, Goodfellow M (eds) *Nucleic Acid Techniques in Bacterial Systematics*. New York: John Wiley & Sons, 1991;175–201.
- Goodbody-Gringley G, Wetzel DL, Gillon D, et al. Toxicity of Deepwater Horizon source oil and the chemical dispersant, Corexit® 9500, to coral larvae. *PLoS One* 2013;**8**:e45574.
- Green SJ, Prakash O, Gihring TM, et al. Denitrifying bacteria isolated from terrestrial subsurface sediments exposed to mixed-waste contamination. *Appl Environ Microb* 2010;**76**:3244–54.
- Hamdan LJ, Fulmer PA. Effects of Corexit® EC9500A on bacteria from a beach oiled by the Deepwater Horizon spill. *Aquat Microb Ecol* 2011;**63**:101–9.
- Hook SE, Osborn HL. Comparison of toxicity and transcriptomic profiles in a diatom exposed to oil, dispersants, dispersed oil. *Aquat Toxicol* 2012;**124–125**:139–51.
- Itah AY, Essien JP. Growth profile and hydrocarbonoclastic potential of microorganisms isolated from tarballs in the Bight of Bonny, Nigeria. *World J Microb Biot* 2005;**21**:1317–22.
- Jones RM, Hedrich S, Johnson DB. *Acidocella aromatica* sp. nov.: an acidophilic heterotrophic alphaproteobacterium with unusual phenotypic traits. *Extremophiles* 2013;**17**:841–50.
- Lee KW, Shim WJ, Yim UH, et al. Acute and chronic toxicity study of the water accommodated fraction (WAF), chemically enhanced WAF (CEWAF) of crude oil plus dispersant in the rock pool copepod *Tigriopus japonicus*. *Chemosphere* 2013;**92**:1161–8.
- Li A, Qu Y, Zhou J, et al. Characterization of a newly isolated biphenyl-degrading bacterium, *Dyella ginsengisoli* LA-4. *Appl Biochem Biotech* 2009a;**159**:687–95.
- Li A, Qu Y, Zhou J, et al. Isolation and characteristics of a novel biphenyl-degrading bacterial strain, *Dyella ginsengisoli* LA-4. *J Environ Sci* 2009b;**21**:211–7.
- Mullen MP, Berry DP, Howard DJ, et al. Single nucleotide polymorphisms in the insulin-like growth factor 1 (IGF-1) gene are associated with performance in Holstein-Friesian dairy cattle. *Front Genet* 2011;**2**:1–9.
- Murakami S, Hayashi T, Maeda T, et al. Cloning and functional analysis of aniline dioxygenase gene cluster, from *Frateria* species ANA-18, that metabolizes aniline via an ortho-cleavage pathway of catechol. *Biosci Biotech Bioch* 2003;**67**:2351–8.
- Muyzer G, De Waal EC, Uitterlinden AG. Profiling of complex microbial populations by denaturing gradient gel electrophoresis analysis of polymerase chain reaction-amplified genes coding for 16S rRNA. *Appl Environ Microb* 1993;**59**:695–700.
- Ortmann AC, Anders J, Shelton N, et al. Dispersed oil disrupts microbial pathways in pelagic food webs. *PLoS One* 2012;**7**:e42548.
- Paixão DAA, Dimitrov MR, Pereira RM, et al. Molecular analysis of the bacterial diversity in a specialized consortium for diesel oil degradation. *Rev Bras Cienc Solo* 2010;**34**:773–81.
- Pietroski JP, White JR, DeLaune RD. Effects of dispersant used for oil spill remediation on nitrogen cycling in Louisiana coastal salt marsh soil. *Chemosphere* 2015;**119**:562–7.
- Place B, Anderson B, Mekebi A, et al. A role for analytical chemistry in advancing our understanding of the occurrence, fate, and effects of Corexit oil dispersants. *Environ Sci Technol* 2010;**44**:6016–8.
- Prakash O, Green SJ, Jasrotia P, et al. *Rhodanobacter denitrificans* sp. nov., isolated from nitrate-rich zones of a contaminated aquifer. *Int J Syst Evol Micr* 2012;**62**:2457–62.
- Prince RC, Butler JD. A protocol for assessing the effectiveness of oil spill dispersants in stimulating the biodegradation of oil. *Environ Sci Pollut R* 2013;**21**:9506–10.

- Prince RC, Lessard RR, Clark JR. Bioremediation of marine oil spills. *Oil Gas Sci Technol* 2003;**58**:463–8.
- Ramachandran SD, Hodson PV, Khan CW, et al. Oil dispersant increases PAH uptake by fish exposed to crude oil. *Ecotox Environ Safe* 2004;**59**:300–8.
- Ramachandran SD, Swezey MJ, Hodson PV, et al. Influence of salinity and fish species on PAH uptake from dispersed crude oil. *Mar Pollut Bull* 2006;**52**:1182–9.
- Rial D, Vazquez JA, Murado MA. Toxicity of spill-treating agents and oil to sea urchin embryos. *Sci Total Environ* 2014;**472**:302–8.
- Rico-Martinez R, Snell TW, Shearer TL. Synergistic toxicity of Macondo crude oil plus dispersant Corexit 9500A® to the *Brachionusplicatilis* species complex (Rotifera). *Environ Pollut* 2013;**173**:5–10.
- Schleheck D, Knepper TP, Eichhorn P, et al. *Parvibaculum lavamentivorans* DS-1^T degrades centrally substituted congeners of commercial linear alkylbenzenesulfonate to sulfophenyl carboxylates and sulfophenyldicarboxylates. *Appl Environ Microb* 2007;**73**:4725–32.
- Schleheck D, Tindall BJ, Rosselló-Mora R, et al. *Parvibaculum lavamentivorans* gen. nov., sp. nov., a novel heterotroph that initiates catabolism of linear alkylbenzenesulfonate. *Int J Syst Evol Micr* 2004;**54**:1489–97.
- Singer MM, George S, Jacobson S, et al. Comparison of acute aquatic effects of the oil dispersant Corexit 9500A with those of other Corexit series dispersants. *Ecotox Environ Safe* 1996;**35**:183–9.
- Singer MM, George S, Lee I, et al. Effects of dispersant treatment on the acute aquatic toxicity of petroleum hydrocarbons. *Arch Environ Con Tox* 1998;**34**:177–87.
- Varadaraj R, Robbins ML, Bock J, et al. Dispersion and biodegradation of oil spills on water. In: *Proceeding of the 1995 Oil spill Conference*. American Petroleum Institute, Washington, DC, USA, 1995;101–106.
- Wang L, Wang W, Lai Q, et al. Gene diversity of CYP153A and AlkB alkane hydroxylases in oil-degrading bacteria isolated from the Atlantic Ocean. *Environ Microbiol* 2010;**12**:1230–42.
- Wang X, Jin D, Zhou L, et al. Draft genome sequence of *Advenella kashmirensis* strain W13003, a polycyclic aromatic hydrocarbon-degrading bacterium. *Genome Announc* 2014;**2**:e00003–14.
- Zemb O, Lee M, Gutierrez-Zamora ML, et al. Improvement of RNA-SIP by pyrosequencing to identify putative 4-n-nonylphenol degraders in activated sludge. *Water Res* 2012;**46**:601–10.
- Zhang Z, Zheng G, Lo IM. Enhancement of nitrate-induced bioremediation in marine sediments contaminated with petroleum hydrocarbons by using microemulsions. *Environ Sci Pollut R* 2015;**22**:8296–306.
- Zhou W, He D, Li X, et al. Isolation and characterization of naphthalene-degrading strains, *Pseudomonas* sp. CZ2 and CZ5. *Afr J Microbiol Res* 2013;**7**:13–9.